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Pollution, sources, and risks of heavy metals in coastal waters of China

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ABSTRACT

Comprehensive information on heavy metals in coastal waters at national scale of China is limited. Therefore, this study investigated the distribution, pollution, and ecological-health risks of heavy metals in coastal waters along 18,000 km coastline of China. Total 13 target heavy metals in coastal waters along coastline of China showed drastic spatial variations with average concentrations ranging from .14 (Cd) to 136.26 (Cu) μ g/L. Cu was the dominant heavy metal with the maximal concentration of 1485.92 μ g/L. Three methods including heavy metal pollution index (HPI), Nemerow index (NI), and contamination degree (CD) were adopted to explore heavy metal pollution. HPI obtained the worst-case evaluation results to illustrate that heavy pollution occurred at over 50% of sampling sites. Anthropogenic sources were the main sources of heavy metals in the coastal waters. Approximately 28.13% and 9.38% of sampling sites illustrated considerable and very high ecological risks, respectively. Metals including Cu, As, and Hg were the main pollution and risk contributors. Heavy metals in coastal waters posed high cancer risks and unacceptable non-cancer risks to both adults and children. Therefore, effective control of heavy metals is necessary for regional sustainability and well-beings of residents in coastal regions of China.

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heavy metal; coastal regions; ecological risks; health risks; source apportionment

Introduction

Coastal regions, possessing several geomorphological features accompanying with specific dynamics (Carapuço *et al.* 2016), play critical functions for human society and ecosystems (Lewison *et al.* 2016). Approximately 40% of global population lived in coastal regions in 2005 (Agardy and Alder 2005), and same in China until now (Meng *et al.* 2017). Covering 13% of the total landmass in China (Meng *et al.* 2017), coastal regions are disturbed by extensive anthropogenic activities in China. Therefore, pollution

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including heavy metal pollution in these regions has been frequently reported (Feng et al. 2011; Wang et al. 2013; Xu et al. 2017).

Heavy metals, referred to metals or metalloids with a density larger than 5 g/cm^3 (Oves *et al.* 2012), have attracted global attention due to inherent toxicity, nondegradability, persistence, and frequent detection in different matrices (Ranjbar Jafarabadi *et al.* 2017; Toro *et al.* 2016; Wang *et al.* 2013; Zhang *et al.* 2017). Heavy metals in coastal water significantly affect water quality and aquatic ecosystem if their concentrations exceed certain thresholds (Tueros *et al.* 2008). Moreover, heavy metals in aquatic system directly or indirectly play critical role in human health owing to their bioaccumulation and biomagnifications through food chain and carcinogenic effects (Zhang *et al.* 2017). Therefore, it is of urgent importance to discuss issues on heavy metals in coastal waters.

China has one of the longest coastlines in the world and stretches from Liaoning $(53^{\circ}30' \text{ N})$ in the north to Guangxi $(18^{\circ}15' \text{ N})$ in the south (Degger *et al.* 2016). Although, pollution in some local coastal regions of China including Bohai Bay, Laizhou Bay, and South China has become research hotspot (Feng *et al.* 2011; Wang *et al.* 2013; Xu *et al.* 2017), comprehensive information on pollution in large-scale region is still limited because most of research focused on local/small-scope areas. It is necessary to obtain comprehensive and thorough information on heavy metal pollution in coastal waters along Chinese coastline to provide basis for making environmental protection policies.

Therefore, this study collected coastal water samples to analyze the concentrations of typical heavy metals along typical coastline of China. The objectives of this study are to determine the distribution of heavy metals in coastal water and discuss the potential ecological-health risks posed by heavy metals in coastal waters. The final aim of this study is to address research basis for sustainable development in coastal regions of China and other countries.

Materials and methods

Study area, sampling, and analysis

Field sampling was carried out during November of 2017. Coastal water samples were collected from four representative coastal regions in China including Yellow Sea Area (Y1-Y11), Bohai Area (B1-B8), East China Sea Area (E1-E6), and South China Sea Area (S1-S7; Figure 1). Sampling locations included different functional zones such as the maricultural zones, bathing beaches, estuaries, and ports.

Surface coastal water samples were collected from .5 m below the water surface using 1 L pre-cleaned amber glass sample bottles. The water samples were filtered using .45 μ m membrane filters (Pall Life Sciences, Ann Arbor, MI, USA) before analysis. Water samples were analyzed using an Agilent7900 inductively coupled plasma mass spectrometry (ICP-MS, Agilent Inc, USA). Concentrations of 13 typical heavy metals including antimony (Sb), arsenic (As), cadmium (Cd), chromium (Cr), cobalt (Co), copper (Cu), vanadium (V), lead (Pb), mercury (Hg), molybdenum (Mo), nickel (Ni), tin (Sn), and zinc (Zn) were determined. Method detection limit (MDL) for As/Cd/Cr/Co/Cu/Sb/V/Pb/Hg/Mo/Ni/Sn/Zn was 164/16/96/12/132/14/29/17/8/29/36/17/336 ng/L. Two purewater samples spiked with standard mixture solution of target heavy metals were

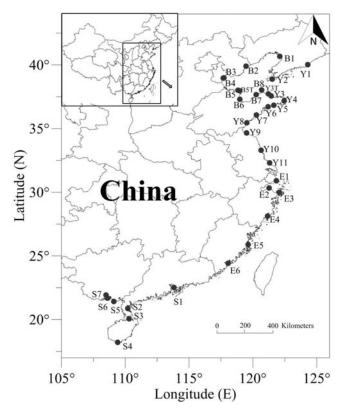


Figure 1. Location of sampling sites.

inserted for analysis series of every 5 real water samples to ensure quality control. The recoveries of target heavy metals ranged from 92.4% to 114.3% to indicate that the ICP-MS analysis was satisfactory for requirement of data quality.

Salinity of each water sample was determined by a portable refractometer (LH-Y100, Lohand Biological, China). Concentrations of total nitrogen (TN) and total phosphorus (TP) in coastal water samples were analyzed using a continuous flow analyzer (AutoAnalyzer III, Seal, Germany). A total organic carbon analyzer (TOC-VCPH, Shimadu, Japan) was used to measure total organic carbon (TOC).

The comparison on concentrations of heavy metals at different sites was evaluated using analysis of variance (ANOVA) by calculating the least significant differences (LSD, p = .01).

Correlation analysis and positive matrix factorization model

Pearson correlation coefficients were employed to determine the relationship between water quality parameters and heavy metals. Correlation analysis was performed using SPSS 19.0 (IBM, NY, USA).

Positive matrix factorization (PMF) was used to determine the possible sources of heavy metals in coastal waters. The detailed information on PMF was provided by USEPA (2014).

Pollution assessment of heavy metals in coastal waters

The Nemerow index (*NI*), heavy metal pollution index (*HPI*), and contamination degree (*CD*) were used to evaluate the pollution of heavy metals in coastal waters along Chinese coastline.

NI was calculated according to the following equation (Liu *et al.* 2015; Vu *et al.* 2017):

$$NI = \sqrt{\frac{\left(\frac{C_i}{S_i}\right)^2_{\text{mean}} + \left(\frac{C_i}{S_i}\right)^2_{\text{max}}}{2}}$$

where C_i and S_i refer to the measured concentration (μ g/L) and assessment standard concentration (μ g/L) of the target heavy metal in coastal water; $\left(\frac{C_i}{S_i}\right)_{\text{mean}}$ and $\left(\frac{C_i}{S_i}\right)_{\text{max}}$ stand for the average value and maximal value of $\left(\frac{C_i}{S_i}\right)$ among all target heavy metals. Values of S_i used in this study were set as the Class II levels of "Sea water quality standard of China (GB 3097-1997)" which targeted at aquaculture zones, bathing beaches, sporting, or entertainment zones, and industrial water supply zones relating with human consumption. Four pollution categories were given including insignificant pollution (Class I, NI < 1), slight pollution (Class II, $1 \le NI < 2.5$), moderate pollution (Class III, $2.5 \le NI < 7$), and heavy pollution (Class IV, $NI \ge 7$) according to evaluation criterion (Liu *et al.* 2015; Vu *et al.* 2017).

HPI model was given by the following equation (Qu et al. 2018):

$$HPI = \frac{\sum_{i=1}^{n} \left(\frac{C_i}{S_i} \times 100\right) \times \left(\frac{k}{S_i}\right)}{\sum_{i=1}^{n} \frac{k}{S_i}}$$

where *k* is a proportionality constant and set as 1 (Qu *et al.* 2018) and *n* represents the numbers of target heavy metals. Comprehensively comparing the current criteria (Edet and Offiong 2002; Qu *et al.* 2018), this study adopted the following evaluation categories: low pollution (Classes I, HPI < 15), moderate pollution (Class II, $15 \le HPI \le 30$), moderate-to-heavy pollution (Class III, $30 < HPI \le 100$), and heavy pollution (Class IV, HPI > 100).

The following equation showed the calculation of CD (Sharifi et al. 2016):

$$CD = \sum_{i=1}^{n} \frac{C_i}{S_i}$$

Sharifi *et al.* (2016) also provided the evaluation criterion: low pollution (Classes I, CD < 6), moderate pollution (Class II, $6 \le CD < 12$), considerable pollution (Class III, $12 \le CD < 24$), and very heavy pollution (Class IV, $CD \ge 24$).

Ecological risk assessment of heavy metals in coastal waters

The potential ecological risks posed by heavy metals in coastal waters were evaluated using the following ecological risk index (*ERI*) model (Sharifi *et al.* 2016):

$$ERI = \sum_{i=1}^{n} T_i \times \frac{C_i}{S_i}$$

where T_i is the potential ecological risk coefficient of a target heavy metal. The values of T_i for all target heavy metals were selected from the literatures (Hakanson 1980; Sharifi *et al.* 2016). Sharifi *et al.* (2016) listed the ranking criterion for ecological risks: low risk (*ERI* < 110), moderate risk ($110 \le ERI < 200$), considerable risk ($200 \le ERI < 400$), and very high risk (*ERI* ≥ 400).

Heath risk assessment of heavy metals in coastal waters

The health risks were evaluated using cancer risks and hazard quotients by the following equations (USEPA 2004):

$$CR(individual) = DAD \times \frac{SF_{O}}{GIABS} = DA_{event} \times \frac{EV \times ED \times EF \times SA}{BW \times AT} \times \frac{SF_{O}}{GIABS}$$
$$= K_{P} \times C_{W} \times t_{event} \times \frac{EV \times ED \times EF \times SA}{BW \times AT} \times \frac{SF_{O}}{GIABS}$$

$$HQ(\text{individual}) = DAD \times \frac{1}{RfD_{O} \times GIABS} = DA_{\text{event}}$$
$$\times \frac{EV \times ED \times EF \times SA}{BW \times AT} \times \frac{1}{RfD_{O} \times GIABS}$$
$$= K_{P} \times C_{W} \times t_{\text{event}} \times \frac{EV \times ED \times EF \times SA}{BW \times AT} \times \frac{1}{RfD_{O} \times GIABS}$$
$$CR = \sum CR(\text{individual})$$

$$HQ = \sum HQ(individual)$$

where DAD and DA_{event} are absorbed dose by dermal contact and absorbed dose of each event, respectively; SF_O represents oral slope factor; GIABS is the fraction of chemical absorbed in gastrointestinal tract; RfD_O refers to oral reference dose; EV, ED, and EF represent the event frequency, exposure duration, and exposure frequency, respectively; SA refers to skin surface area; BW means body weight; AT is average lifespan; K_P refers to dermal permeability coefficient of pollutant; C_W is concentration of pollutant in coastal water; t_{event} represents event duration. The values of parameters were obtained from the references (Akhbarizadeh *et al.* 2016; IRIS 2018; Ranjbar Jafarabadi *et al.* 2017; Sarria-Villa *et al.* 2016; USEPA 2004, 2016). The values of parameters were listed in Table 1. Based on ranking criterion (Ge *et al.* 2013), cancer risks were categorized into five levels including very low ($CR \le 1 \times 10^{-6}$), low ($1 \times 10^{-6} < CR \le 1 \times 10^{-4}$), moderate ($1 \times 10^{-4} < CR \le 1 \times 10^{-3}$), high ($1 \times 10^{-3} < CR \le .1$), and very high (CR > .1). Unacceptable threshold of non-cancer risks was set as 1 (Qu *et al.* 2018; Zhang *et al.* 2017).

Table 1. Parameters employed for human health risk assessment.	or human health r	isk assessment.		
		Va	Value	
Parameter	Unit	Children	Adults	References
EV (event frequency)	event/day	1	Ţ	Akhbarizadeh et al. (2016); USEPA (2004)
ED (exposure duration)	year	9	26	USEPA (2004, 2016)
EF (exposure frequency)	days/year	150	150	Akhbarizadeh et al. (2016)
SA (skin surface area)	cm ²	6600	20,000	Akhbarizadeh et al. (2016); USEPA (2004)
<i>BW</i> (body weight)	kg	15	60	USEPA (2004); Zhang et al. (2017)
AT (average lifespan)	days	2190 (for CR) 2190 (for HQ)	25,550 (for CR) 9490 (for HQ)	Ranjbar Jafarabadi et al. (2017); Sarria-Villa et al.
				(2016); USEPA (2004)
SF_O (oral slope factor)	(kg·day)/mg	Chemical-specific	Chemical-specific	IRIS (2018); USEPA (2004)
RfD_O (oral reference dose)	mg/(kg·day)	Chemical-specific	Chemical-specific	IRIS (2018); USEPA (2004)
GIABS (absorption fraction in GIT)	unitless	Chemical-specific	Chemical-specific	USEPA (2016)
$K_{ ho}$ (dermal permeability coefficient)	cm/h	Chemical-specific	Chemical-specific	USEPA (2004)
t _{event} (event duration)	h/event	.58	.58	USEPA (2004)
C_W (pollutant concentration)	mg/L	Measured	Measured	This study

Note: GIT refers to gastrointestinal tract; CR and HQ refer to cancer risk and hazard quotient, respectively.

Results and discussion

Occurrence and distribution of heavy metals in coastal waters

Statistical summary on concentrations of heavy metals in coastal waters was shown in Table 2. The average concentrations of heavy metals ranged from .14 (Cd) to 136.26 (Cu) μ g/L. Significant differences in concentrations of heavy metals were proved by ANOVA results. Cu was the dominant heavy metal in coastal waters with the concentrations ranging from 6.24 (at S2) to 1485.92 (at B5) μ g/L, illustrating the most significant spatial variation. As surprisingly showed high concentrations varying from 4.20 (at S2) to 112.64 (at B2) μ g/L with the average value of 78.08 μ g/L. Zn and V also showed the relatively high average concentrations. The minimal/maximal concentrations of Zn and V occurred at Y5/Y3 and S2/B5, respectively. Concentrations of Mo varied from 1.91 (at S2) to 14.35 (at B5) μ g/L, showing the relatively low fluctuation in contrast with Cu, Ni, Zn, Co, V, Cd, Sn, Sb, Hg, and Pb. Except E4 and S2 at which Ni was not detected, concentrations of Ni ranged from 2.48 (Y1) to 46.68 (at B5) μ g/L. Except Y4 and Y6 at which Pb was not detected, concentrations of Pb were in the range of .02 (at Y3)-5.67 (at Y5) μ g/L. Concentrations of Cr were in the range of 1.36 (at S7)-4.68 (at B5) μ g/L, also exhibiting relatively low fluctuation similar to Mo. Concentrations of Hg ranged from .006 (at E3) to 1.68 (at Y8) μ g/L at sampling sites except Y9, Y10, E5, E6, S1, and S3-S7. The minimal and maximal concentrations of Cd occurred at B5 and E4, respectively. In summary, distribution of heavy metals exhibited drastic spatial variations.

The maximal average concentration of many target heavy metals (Cr, Co, Ni, Cu, As, Mo, Cd, Sn, Sb) in coastal waters occurred in Bohai Area. The maximal average concentration of V, Zn, and Hg in coastal waters occurred in Yellow Sea Area while that of Pb occurred in South China Sea Area. Average concentration of Cr, As, and Mo in coastal waters followed the order of Bohai Area > Yellow Sea Area > East China Sea Area > South China Sea Area while that of Co, Ni, and Cu followed the order of Bohai Area > Yellow Sea Area. Average concentration of V and Zn in coastal waters followed the order of Yellow Sea Area > Bohai Area > South China Sea Area > East China Sea Area > Bohai Area > South China Sea Area > East China Sea Area > Bohai Area > South China Sea Area > East China Sea Area > Bohai Area > South China Sea Area > East China Sea Area > Heat Area > South China Sea Area > East China Sea Area > Bohai Area > South China Sea Area > East China Sea Area > South China Sea Area > East China Sea Area > South China Sea Area > East China Sea Area > South China Sea Area > East China Sea Area > South China Sea Area > East China Sea Area > South China Sea Area > East China Sea Area > South China Sea Area > Yellow Sea Area > Yellow Sea Area > Yellow Sea Area > Kea > South China Sea Area > Yellow Sea Area > Yellow Sea Area.

Concentrations of several heavy metals including Cr, Cd, Ni, Zn, and Pb in coastal waters of this study were lower than those in surface water near Yeongil Bay, Korea

Table 2. Statistical summary on salinity, TOC, TN, TP, and heavy metal concentrations in coastal waters (μ g/L for all heavy metals; $%_{00}$ for salinity; mg/L for TOC, TN, and TP).

	V	Cr	Co	Ni	Cu	Zn	As	Мо	Cd	Sn	Sb	Hg	Pb	salinity	TOC	ΤN	TP
Mean	29.44	2.55	.92	13.59	136.26	32.42	78.08	10.09	.14	.39	1.38	.56	2.08	26	6.32	1.93	.05
SE	2.86	.14	.07	1.39	45.52	2.73	3.82	.59	.01	.06	.13	.10	.26	2	.98	.15	.01
Median	25.55	2.27	.89	13.35	60.52	27.38	82.59	11.16	.13	.29	1.22	.44	1.76	30	4.50	1.97	.04
Minimum	4.09	1.36	.21	BDL	6.24	18.30	4.20	1.91	.02	.03	.43	BDL	BDL	2	2.47	.56	.01
Maximum	59.47	4.68	1.68	46.68	1485.92	88.40	112.64	14.35	.30	1.55	4.35	1.68	5.67	35	29.42	3.87	.17

Note: TOC, TN, and TP refer to total organic carbon, total nitrogen, and total phosphorus; SE means standard error; BDL means below the detection limit.

(Kim *et al.* 2014). Concentrations of Cd and Co in coastal waters of China (this study) were lower than those in Sagami River and Bay of Japan (Takata *et al.* 2012). Concentrations of Cd, Cr, Ni, and Pb in coastal waters of SW Gulf of California coast were higher than those of this study (Jonathan *et al.* 2017) while concentrations of Pb and Cd in coastal waters of this study were comparable with those in coastal regions of South-Western Sardinia, Italy (Cherchi *et al.* 2009).

Heavy metal pollution along Chinese coastline

Pollution of individual heavy metals including Cr, Ni, Cu, Zn, As, Cd, Hg, and Pb was evaluated by comparing corresponding Class II standards of "Sea water quality standard of China (GB 3097-1997)" which only listed these metals. Concentrations of Cd and Cr in coastal waters were much less than Class II levels of 5 (Cd) and 100 (Cr) μ g/L, respectively. Concentrations of Pb at B5 and S1 exceeded its standard (5 μ g/L) while those at the remaining sites were less than quality standard. Concentrations of Zn at 3 sites (B5, Y3, and S1) exceeded its standard (50 μ g/L) while those at the remaining sites were less than quality standard. Concentrations of Zn at 3 sites (B5, Y3, and S1) exceeded its standard (50 μ g/L) while those at the remaining sites were less than quality standard. Cu and As showed significant pollution since concentrations of them at 30 sites exceeded Class II standards. Concentrations of Ni at 25 sites exceeded its standard (10 μ g/L) while those of Hg at 16 sites exceeded the Class II standard (.2 μ g/L). Heavy metals including Cu, As, Ni, Hg, Zn, and Pb generally caused pollution in coastal waters to different extents.

Three methods were used to evaluate the comprehensive pollution caused by Cr, Ni, Cu, Zn, As, Cd, Hg, and Pb since only these 8 metals possessed assessment standard concentrations (Figure 2). *NI* values ranged from 1.91 (at S2) to 106.00 (at B5). Based on *NI* ranking criterion, total 6, 14, and 12 out of 32 sampling sites exhibited slight, moderate, and heavy pollution, respectively. *NI* used average and maximal values of $\frac{C_i}{S_i}$ for evaluation so that heavy metals possessing the highest $\frac{C_i}{S_i}$ value would become the dominant pollution contributor. Metalloid As served as the predominant contributor of pollution at sampling sites Y1, Y10, E1, E2, E3, and S7. Hg was the main contributor of pollution at the remaining sites, followed by As and Hg.

HPI values of sampling sites varied from 7.39 (at Y10) to 787.98 (at Y8). Approximately 25%, 18.75%, 3.13%, and 53.13% of sampling sites showed low, moderate, moderate-to-heavy, and heavy pollution, respectively. Cu served as the predominant pollution contributor of 13 sampling sites (B5, B7, Y9, Y10, E3, E5, E6, S3, S4, S5, S6, and S7) while Hg was the main pollution contributor for the rest of sampling sites. Heavy metal pollution assessed by *HPI* was more serious than that evaluated by *NI*.

CD values ranged from 4.8 (at S2) to 158.78 (at B5). Based on *CD* ranking criterion, approximately 2, 15, 5, and 10 out of 32 sampling sites illustrated low, moderate, considerable (heavy), and very heavy pollution, respectively. Cu was the main pollution contributor for most of sampling sites. Heavy metal pollution assessed by *CD* was also more serious than that evaluated by *NI*.

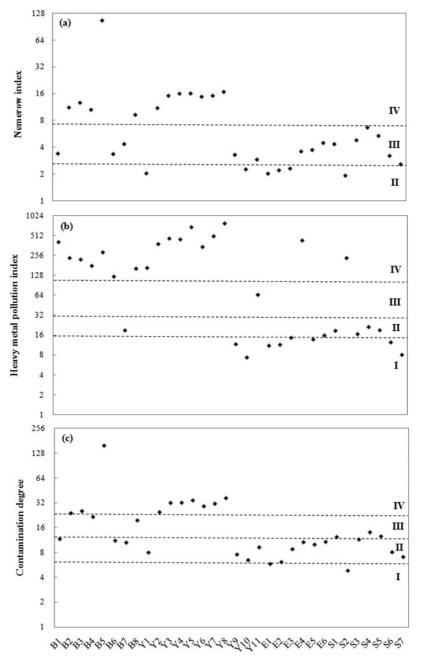


Figure 2. Heavy metal pollution of coastal waters using different evaluation methods. Classes II, III, and IV of Nemerow index represent slight, moderate, and heavy pollution, respectively; Classes I, II, III, and IV of heavy metal pollution index refer to low, moderate, moderate-to-high, and heavy pollution, respectively; Classes I, II, III, and IV of contamination degree stand for low, moderate, considerable, and very heavy pollution, respectively.

In general, Yellow Sea Area and Bohai Area showed more serious heavy metal pollution than East China Sea Area and South China Sea Area. The most serious heavy metal pollution occurred in the region near B5. Northern Chinese coastline needs urgent and effective heavy metal pollution control and management.

Possible sources of heavy metals in coastal waters

Pearson correlation coefficients were used to analyze the potential relationship among heavy metals and water quality parameters (Table 3). Significantly positive relationship existed between every two metals of V, Cr, Co, Ni, Cu, and Mo, indicating that these metals might originate from the same source. Sn was significantly related with V and Sb at p < .01 while Pb was significantly related with Hg and Sb at p < .01. Cu was positively related with Zn and negatively related with Cd at p < .05while Hg was also related with Cr, Co, and Sn at p < .05. Salinity was significantly related with V, Cr, Co, Ni, Mo, and Sn while TOC was significantly related with Co, Mo, Sn, Sb, and Pb. TN was significantly related with As, Sn, and Hg while TP was significantly negatively related with Hg. Complex relationship between heavy metals illustrated that multiple sources might affect the distribution of heavy metals in coastal waters.

Three possible sources were determined based on PMF analysis (Figure 3). Metals including V, Co, Ni, Cu, As, Mo, and Cd were mainly controlled by Factor 3 that might originate from discharge of industrial wastewater and sewage, agricultural runoff, and mining inputs. Pb, Zn, and Sn were mainly from industrial wastewater and traffic source (Factor 2) since these three metals frequently occurred in industrial wastewater and Pb were also an important component of fuel. Hg was mainly controlled by Factor 1 that might be from combustion of fossil fuels and reverine inputs. Heavy metals in coastal waters often concurrently originated from three possible sources, suggesting the source complexity and challenge of effective pollution control. Three possible sources were mainly triggered by the anthropogenic activities, illustrating that rational industry planning is critical for sustainability of the coastal regions.

Potential ecological risks of heavy metals in coastal waters

Potential ecological risks of heavy metals were evaluated using ecological risk index based on ranking criterion (Figure 4). Approximately 15, 5, 9, and 3 out of 32 sampling sites illustrated low, moderate, considerable, and very high risk, respectively. Cu, Hg, and As accounted for most of the ecological risks, serving as the dominant risk contributors for 14, 13, and 5 sampling sites, respectively. Sites B5, Y5, and Y8 exhibited very high ecological risks, which corresponded with very high heavy metal pollution at the same site. Ecological risks of heavy metals in 64% of sampling sites in Yellow Sea Area were considerable or very high, more serious than those in the other areas. Accordingly, approximately 50% of sampling sites in Bohai Area showed considerable or very high ecological risks. South China Sea Area showed low (86%) or moderate (14%) ecological risks increased with more strict assessment standard concentrations (S_i). Approximately 6, 7, 4, and 15 sites showed low, moderate, considerable, and very high risk, respectively when Class I standards of "Sea water quality standard of China (GB 3097-1997)" were

Table 3	able 3. Pearson correlation coefficients	correlati	on coeffic		veen tra	ce elemo	ents and	water qu	uality pa	between trace elements and water quality parameters.							
	>	ა	c	İN	J	Zn	As	Mo	Cd	Sn	Sb	Hg	Pb	Salinity	TOC	TN	đ
>	1.000																
ڻ	.798**	1.000															
ĉ	.890**	.733**	1.000														
ïZ	.747**	.716**	.820**	1.000													
Cu	.584**	.628**	.580**	.917**	1.000												
Zn	.227	660.	.226	.426*	.425*	1.000											
As	.234	.072	.093	053	.132	.050	1.000										
Mo	.813**	.608**	.898**	.696**	.429*	.092	.229	1.000									
Cd	047	030	.190	154	369*	029	229	.244	1.000								
Sn	514**	297	277	247	229	102	.026	336	.129	1.000							
Sb	236	029	001	.101	.134	.070	.077	042	.041	.682**	1.000						
Hg	.546**	.440*	.459*	.087	057	070	173	.355	020	444*	413	1.000					
Pb	231	.074	176	.303	.343	.303	079	307	147	.363*	.476**	608**	1.000				
Salinity	.797**	.529**	.799**	.572**	.299	.045	.119	.888**	.277	461**	296	.387	345	1.000			
TOC	484**	226	433*	243	131	035	179	484**	109	.397*	.464**	319	.467**	693**	1.000		
NT	222	146	229	261	184	246	.385*	.011	.028	.422*	.265	429*	.084	015	078	1.000	
₽	333	322	198	185	150	026	.205	010	.327	.345	.311	538**	.007	050	063	.543**	1.000
Note: **	Vote: ** means significance level at $p<.01$; *	nificance lev	vel at $p <$		means significance level at $p < .05$.	ce level a	at <i>p</i> < .05.										

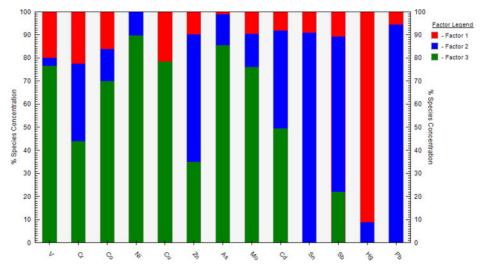


Figure 3. Factor fingerprints of heavy metals in coastal waters along Chinese coastline.

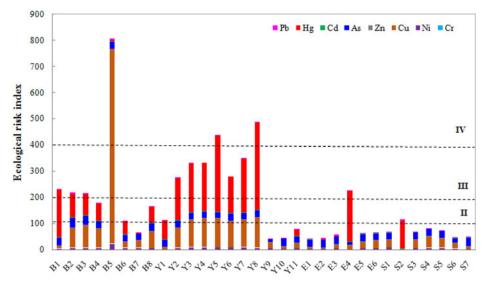


Figure 4. Potential ecological risks posed by heavy metals in coastal waters. Classes I, II, III, and IV refer to low risk, moderate risk, considerable risk, and very high risk, respectively.

used as S_i values to evaluate the risks, more serious than those assessed by Class II standards. Ecological risk index method is usually employed to determine the potential ecological risks posed by heavy metals in soils, sediments, and dusts (Barać *et al.* 2016; Hakima *et al.* 2017; Trujillo-González *et al.* 2016) while its application in risk assessment on heavy metals in water especially coastal waters is limited. Therefore, this study provided new insight on ecological risks posed by heavy metals in coastal waters along Chinese coastline. Effective strategies should be proposed to control potential ecological risks of heavy metals in Chinese coastal waters.

Potential health risks of heavy metals in coastal waters

Heavy metals in coastal waters posed potential health risks mainly through dermal contact. Therefore, health risks of heavy metals were evaluated using cancer risks and noncancer risks (Figure 5). Cancer risks were in the range of 8.71×10^{-3} - 8.32×10^{-2} for adults and 3.09×10^{-2} - 2.96×10^{-1} for children, respectively. Heavy metals in coastal waters of the study area exerted high cancer risks for adults at all sampling sites while they posed high and very high cancer risks for children at 50% of sampling sites, respectively. Cancer risks for children were usually several times those for adults, suggesting that children are more susceptible to hurt of heavy metals. V, Cr, As, Ni, and Cu were the main contributors for cancer risks, accounting for 90.98–98.60% of cancer risks.

Hazard quotients were in the range of 6.44-32.12 for adults and 22.90-114.13 for children, respectively. Hazard quotients for adults and children all exceeded the

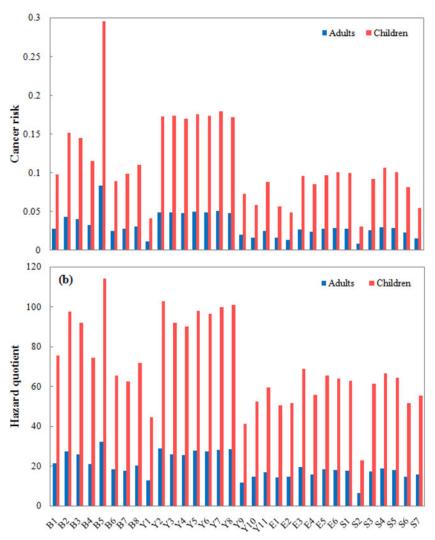


Figure 5. Cancer risks (a) and non-cancer risks (b) posed by heavy metals in coastal waters.

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unacceptable threshold of 1.0, suggesting that non-cancer risks posed by heavy metals in coastal waters were high for human beings, especially children. Elements including As, Cr, V, Hg, and Sb were the main contributors for non-cancer risks, accounting for 93.96–98.88% of non-cancer risks.

Heavy metals in coastal waters along coastline of China exerted high health risks for human beings. Therefore, effective control of heavy metals is necessary for regional sustainability and well-beings of residents.

Conclusions

Total 13 heavy metals were detected in coastal waters along Chinese coastline with significant spatial variations. Cu was the dominant heavy metal with the concentrations ranging from 6.24 to 1485.92 μ g/L. Heavy metal pollution index approach among three assessment methods obtained the worst-case evaluation results, showing that heavy pollution of heavy metals in coastal waters occurred at over 50% of sampling sites. Metals including Cu, As, and Hg were the main pollution contributors. Anthropogenic sources such as industrial wastewater, sewage, traffic, and fuel combustion were the main sources of heavy metals in the coastal waters. Taking Class II levels of "Sea water quality standard of China (GB 3097-1997)" as assessment standards, approximately 46.88%, 15.63%, 28.13%, and 9.38% of sampling sites illustrated low, moderate, considerable, and very high ecological risks, respectively. Cu, Hg, and As also accounted for most of the ecological risks. Heavy metals in coastal waters posed high cancer risks for adults at all sampling sites while they exerted high and very high cancer risks for children at 50% and 50% sampling sites. Heavy metals posed unacceptable non-cancer risks for both adults and children. Therefore, heavy metal pollution control in the coastal regions of China is in urgent need.

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