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Marine Pollution Bulletin



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Redox-dependent phosphorus burial and regeneration in an offshore sulfidic sediment core in North Yellow Sea, China



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ARTICLE INFO	A B S T R A C T
<i>Keywords:</i> P sequential extraction ³¹ P NMR Sulfide Redox Marine sediments	Phosphorus (P) pollution can trigger severe marine eutrophication, which further leads to harmful algal blooms, and a deterioration of sea water quality. The P burial and regeneration in offshore sediments can directly affect the eutrophication levels of estuarine and coastal ecosystems. Although many researches on redox-dependent P burial and regeneration were studied, such process in the presence of silicate is still poorly understood, and the effects of pyrite formation on organic P (OP) burial and regeneration also remain unclear. In this study, a sulfidic sediment core was collected in the offshore of an estuary in the north Yellow Sea, China. Results indicated that indigenous biological input was found to be the primary source of organic matter in upper sediments. The regenerated P under reducing conditions was dominated by labile Fe–P and OP. The PO ₄ ^{3–} released from Fe–P and OP that could be captured by Al/Fe/Mn (oxyhydr) oxides in surface sediments and Ca minerals in deep sediments. Ca–P, Al–P, unreactive Al/Fe-Si-P and some stable metal chelated OP were the main burial P fractions. Sulfate reduction and formation of insoluble metal sulfides including the pyrite promoted OP decomposition by anaerobic decomposi-

tion, removing metal ions from the "metal-OP" chelates and restoring the phosphatase activity.

1. Introduction

Phosphorus (P) is an essential nutrient for organism growth and a limiting factor for marine primary productivity (Smith, 1984). Although P is removed from marine systems by burial in sediment, abundant P accumulated in sediment can be regenerated into the overlying water, contributing to algal blooms (Eijsink et al., 2000). The mechanisms of sedimentary P burial and regeneration are closely related to the sediment redox, P speciation, and sulfidic conditions (Li et al., 2016; Sun et al., 2016; Zhao et al., 2019b).

In upper sediments with oxidizing condition, soluble reactive phosphorus (SRP) from the overlying water and regenerated PO_4^{3-} from the decomposition of organic P (OP) in plant debris are captured by Al/Fe/Mn (oxyhydr) oxides and Ca compounds transforming into Al/Fe/Mn (oxyhydr) oxides-bound P (Al/Fe/Mn-P) and Ca compounds-bound P (Ca–P) (Zhu et al., 2013; Li et al., 2018). In deep sediments with reducing condition, sulfate reduction becomes strong, accompanied by dissolved sulfide and ferrous iron (Fe II) accumulation. The Fe-bound P can be released during ferric iron (Fe III) reductive dissolution, and the generated Fe(II) rapidly reacts with dissolved HS⁻ (Krom and Berner, 1980). However, the concentrations of Si and Al are higher than Fe, Mn

and Ca contents in marine sediment, which can potentially alter the relationship between bulk sediment geochemistry and P burial mechanisms. For instance, there are many stable Al-bound silicate minerals (e.g., kaolinite, feldspar and dickite) in sediments (Story et al., 2010). P can not only be adsorbed on the surfaces of Al and/or Fe bound silicate minerals to form stable P complexes but also be preserved within the silicate minerals by Al–OH and Fe–OH functional groups through ion exchanges (Kasama et al., 2004; Ruttenberg and Sulak, 2011), which enhance capacity for P sequestration. The P within the silicate minerals is called occluded P, which is too stable to be released under natural conditions (Liu et al., 2003). Therefore, to investigate redox-dependent P burial and regeneration mechanism in the presence of Al/Fe-bound silicate minerals will provide new insights in P cycling.

P burial and regeneration is closely related to solid-phase P speciation (Carpenter, 2008; Dijkstra et al., 2018). There are many classical P chemical speciation analysis methods, such as Standards, Measurements and Testing (SMT) Program, Ivanoff organic P (OP) sequential extraction, and solution ³¹P-nuclear magnetic resonance spectroscopy (³¹P NMR) (Ivanoff et al., 1998; Ruban et al., 2001; Cade-Menun and Liu, 2014). The SMT protocol consists of NaOH-extractable inorganic P (Al/Fe/Mn-P, P bound to Al, Fe and Mn oxides and

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https://doi.org/10.1016/j.marpolbul.2019.110582

Received 12 May 2019; Received in revised form 2 September 2019; Accepted 7 September 2019 Available online 26 September 2019

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hydroxides), HCl-soluble inorganic P (Ca-P, P associated with Ca), OP, and total P (TP). The Ivanoff method divides OP into readily labile OP, moderately labile OP and nonlabile OP. Al/Fe/Mn-P and labile OP are easily released and hydrolyzed into soluble reactive PO₄³⁻, while Ca–P and non-labile OP are too reluctant to be easily regenerated (Zhao et al., 2018; Zhu et al., 2018). In addition, molecular structure and chemical state of P can affect its burial and regeneration (Zhu et al., 2015). The solution ³¹P-nuclear magnetic resonance spectroscopy (³¹P NMR) method is a powerful technology in determination for the molecular structure of P compounds. The ³¹P NMR method can identify specific P compounds (e.g., Glucose 6-phosphate, inositol hexaphosphoric acid, monophosphate, DNA, RNA, phospholipids and polyphosphate) in EDTA and NaOH extraction of sediments (Cade-Menun and Preston, 1996). Inositol hexaphosphoric acid and polyphosphate are two kinds of natural metal chelating agents, which can react strongly with metal ions to form insoluble complexes (Turner et al., 2012; Huang et al., 2018). Previous studies indicated that free of glucose 6-phosphate, monophosphate and DNA can be easily hydrolyzed into PO₄³⁻ by phosphatase. However, humic acid and/or metal ions associated with these OP compounds become fairly recalcitrant, which can be stably preserved in sediments (Zhu et al., 2015, 2018). Therefore, an accurate understanding of redox-dependent P burial and regeneration in marine sediments based on multiple methods is essential.

Sulfate reduction can produce toxic free dissolved sulfide (referred to here as HS⁻, dominant form at seawater pH) (Kraal et al., 2017), which can precipitate with divalent metal ions including $Fe^{2+/3+}$ to generate insoluble metallic sulfides in sediments (Ankley, 1996). Therefore, we hypothesized that the formation of insoluble metallic sulfides reduced the efficiency of metal ions combining with OP. The free OP would be easily hydrolyzed into PO_4^{3-} by phosphatase after the metal ions were removed by HS⁻. In addition, OP can be used as electron donor during the sulfate reduction resulting in the decomposition of OP. Therefore, sulfate reduction may improve the risk of regeneration of OP.

The PO_4^{3-} pollution in coastal areas often leads to toxic algal blooms, which can seriously poison marine organisms. The regeneration of internal P accumulated in sediments has become an important source of PO_4^{3-} in coastal waters due to anthropogenic inputs of PO_4^{3-} have gradually been increasingly controlled by strict environmental laws (Li et al., 2018). Therefore, the study of sedimentary P burial and regeneration is essential for controlling P endogenous pollution in marine. The objectives of this study were to (1) identify the potential sources of P and the time of algae bloom by mass ratio of total organic carbon (TOC) to total nitrogen (TN), and carbon and nitrogen isotope ($\delta^{13}C$ and $\delta^{15}N$); (2) investigate redox-dependent P burial and regeneration mechanism based on the various P speciation in the presence of Si and Al composite oxides and (3) recognize the effect of sulfate reduction and pyrite formation on the regeneration of OP.

2. Materials and methods

2.1. Study area and sample collection

The sampling site $(37^{\circ}46.93'N, 121^{\circ}19.39'E)$ was located near Jiahe River estuary in the north Yellow Sea (NYS), China (Fig. 1), which is a typical epicontinental marginal sea with an average water depth of ~38 m (Yang et al., 2010). The Jiahe River is one of the principal rivers along the Shandong Peninsula coast, entering the NYS with high loadings of anthropogenic material input. A sediment core (370 cm deep) was collected using a 400-cm long polycarbonate tube with an inner diameter of 10 cm. The deep sediment core can provide valid results and interpretations because the Si, Fe, Al and Ca are the inherent and extensive elements in earth's crust. Distributions and contents of Si, Al, Fe and Ca do not vary significantly among different sampling areas, and the rules of vertical variation of redox condition and sulfate reduction in sediments are universal in sediments. The spatial variation cannot change the above basic rules of geochemistry. Therefore, the effect mechanism of Si, Fe and Al composite oxides, redox and sulfate reduction on P burial and regeneration can be identified by a sediment core.

The sediment core was sectioned at intervals of 2–20 cm. Segments of the cut sediment were stored in sealed plastic bags filled with N_2 at -20 °C, then transported to laboratory within 3 h. Before analysis, all sediments were lyophilized and homogenized and ground to powder in a sealed N_2 -filled chamber.

2.2. Chemical analysis

Sediment samples were pretreated with 1 M HCl to remove carbonates, and then total organic carbon (TOC) and total nitrogen (TN) were determined using a Vario MACRO cube CNS analyzer (Elmentar, Germany). Subsamples after the decarbonization were selected for the organic carbon and nitrogen isotope (δ^{13} C and δ^{15} N) analyses. δ^{13} C and δ^{15} N analyses were conducted using a Delta ^{plus} XL mass spectrometer connected with a Flash EA elemental analyzer (Thermo Fisher MAT253, USA). δ^{13} C and δ^{15} N, reported in parts per million (‰), were calculated as follows:

 $\delta X (\%) = [(R_{sample}/R_{standard}) - 1] \times 10^3$

where X is ¹³C or ¹⁵N, and R_{sample} and R_{standard} are the isotopic ¹³C/¹²C ratios of the samples and standards, respectively. The carbon standard is Vienna Pee Dee Belemnite and the nitrogen standard is air. The standard error of δ^{13} C and δ^{15} N analysis was below 0.2‰ and 0.4‰, respectively.

The concentrations of Al, Fe, Ca and Mn in sediments were determined using an inductively coupled plasma-optical emission spectroscopy (ICP-OES, Perkin-Elmer Optima 7000 DV, USA) after microwave-acid wet digestion by HNO_3 , HCl, and HF (Sneddon et al., 2006). The Al, Fe and Ca bound silicate minerals in the adequately homogenized samples from -2 cm, -90 cm, -150 cm and -370 cm depths were analyzed with an energy dispersive X-ray micro-analyzer (EDS, EX-350, HORIBA, Japan) at an acceleration voltage 15 kV. The target elements and their composition in mol% were selected and estimated using AZtec software (Dijkstra et al., 2018), which excluded carbon, sulfur, potassium, titanium, copper and chloride (manually inserted in the figures) (Figs. 2 and 3).

Chlorophyll *a* was determined by spectrophotometry at 663 nm and 645 nm after extracting with 80% (vol%) acetone in a dark environment at 4 °C (Arnon, 1949). Measurement of alkaline phosphatase reactivity (APA) was based on the standard *p*NPP (*p*-nitrophenyl phosphate) method using a spectrophotometer at 410 nm (Sayler et al., 1979). The reactive Fe was extracted from the sediments using 1 mol L⁻¹ HCl (Wallmann et al., 1993). Total reactive Fe (Fe _{RT}) and Fe (II) concentrations in the extracts were determined with the colorimetric phenanthroline method (Stookey, 1970). Fe(II) and Fe _{RT} were determined by measuring the absorbance before and after adding hydroxylamine-hydrochloride to reduce agent that can convert any dissolved Fe(III) to Fe(II). The Fe(III) concentration was calculated as the difference between Fe _{RT} and Fe(II).

Reduced inorganic sulfur (RIS) pools were determined using a slight modification of the sequential extraction method of Sheng et al. (2015), as described by Newton et al. (1995) and Hsieh and Shieh (1997). The RIS was divided into acid-volatile sulfide (AVS; 6 mol L⁻¹ HCl, 1 h), chromium-reducible sulfide (CRS or pyrite-S; 15 mL acidic Cr^{2+} , 2 h), and elemental sulfur (ES; 5 mL Cr^{2+} + 20 mL *N*,*N*-dimethylformamide (DMF) + 5 mL concentrated HCl, 1 h). The acid-volatile sulfide gives an estimate of the Fe monosulfide (FeS) (Rickard and Morse, 2005). The sulfide released as H₂S during AVS, CRS and ES extraction was trapped as CuS in CuCl₂ traps (Newton et al., 1995). The sulfide content of the traps was determined by iodometric titration (Kramar et al., 2014). Detailed analysis steps of the above physicochemical parameters and data quality control (QC) and quality assurance (QA) are presented in the Supporting information.

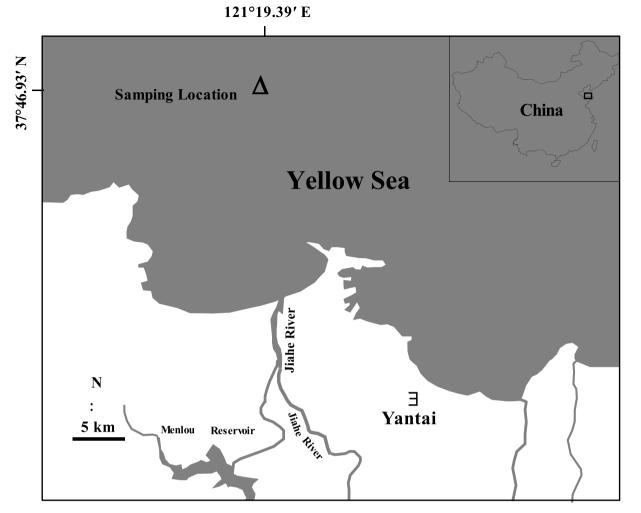
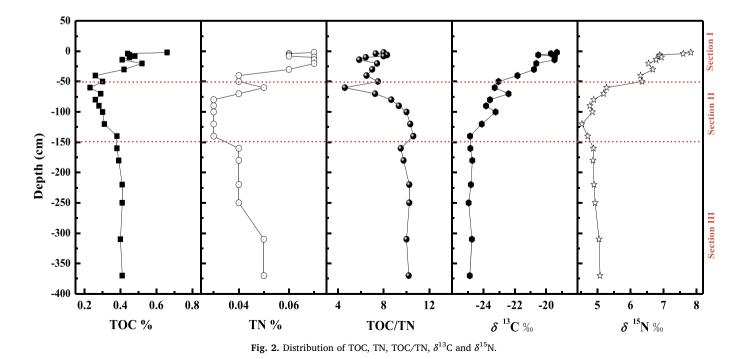


Fig. 1. Sampling location of the core sediment in the North Yellow Sea.



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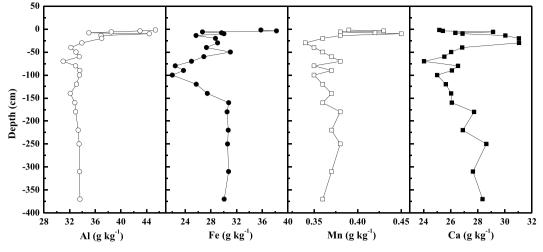


Fig. 3. Distribution of Al, Fe, Mn and Ca in the sediment core.

2.3. Phosphorus sequential extraction

The chemical sequential fractionation of IP in sediments was determined using the Standards, Measurements and Testing (SMT) protocol (Ruban et al., 2001). The extraction of P in the protocol was divided into Al/Fe/Mn-P, Ca–P, OP and TP. The sedimentary OP was sequentially extracted with the Ivanoff method (Ivanoff et al., 1998; Zhao et al., 2018). OP in sediments was divided into five fractions: NaHCO₃-extractable OP (NaHCO₃-OP), HCl-soluble OP (HCl-OP), OP in the form of fulvic acid (Flv-OP), humic acid (Hum-OP), and residual OP (H₂SO₄-OP). Additionally, NaHCO₃-OP was defined as readily labile OP, and Hum-OP were treated as nonlabile OP. Detailed analysis steps of the P chemical fractionation and data QC/QA are presented in the Supporting information.

2.4. NaOH-EDTA extraction and ³¹P NMR analysis

Using a widely applied extraction procedure, each freeze-dried sample was mixed with NaOH-EDTA solution at a 1: 10 w/v ratio, shaken for 16 h at 20 °C, and centrifuged (Turner et al., 2005). Briefly, approximately 3.0 g of lyophilized sample was added to 30 mL NaOH-EDTA solution (0.25 mol L^{-1} NaOH and 50 mmol L^{-1} Na₂EDTA). The supernatant was collected after centrifugation at 12,000 rpm for 30 min. The supernatant was collected for TP and IP determination. TP in extracts was digested with $K_2S_2O_8 + H_2SO_4$ in an autoclave at 121 °C for 30 min (Zhao et al., 2018). IP in extracts was directly determined without the digestion. The extracted OP was calculated as the difference between TP and IP. The recoveries of TP and OP in extracts were defined as the ratio of TP in extracts to TP in sediments, and extracted OP to OP in sediments, respectively (Zhang et al., 2008).

The remaining extracts were lyophilized for 24–48 h to obtain the dry extracts. Approximately 200 mg lyophilized powder was re-dissolved in 0.6 mL D₂O and 0.1 mL 10 mol L⁻¹ NaOH, vortexed, and centrifuged at 13,000 rpm for 20 min at 4 °C. The supernatant was transferred to a 5-mm NMR tube for analysis. The solution ³¹P NMR spectra were obtained with an Avance III Bruker 500 MHz spectrometer (Bruker BioSpin GmbH, Germany) at 202.47 MHz with a 30° pulse, 0.58 s acquisition, a 2.0 s T₁ delay, and 20 °C temperature. The spectra were collected based on approximately 10,000 scans. The free induction decay was not truncated in the sample analysis. The pulse angle, acquisition time, T₁ delay, temperature and scan times were selected to obtain quantitative spectra based on a previous report (Turner et al., 2012).

The signal of P species spectrum was identified and quantified based on spiking experiments and the chemical shifts relative to documented data, with the orthophosphate peak in all spectra standardized to 6 ppm (Cade-Menun, 2005, 2015). Four sediments (-2 cm, -90 cm, -150 cmand -370 cm) were selected to conduct the standard addition experiment. The same NaOH-EDTA extractions were spiked with β -glycerophosphate disodium salt hydrate (Sigma number G5422), DL- α -glycerol phosphate (Sigma number 17766), adenosine 5' monophosphate disodium salt (Sigma number 01930), phosphocholine chloride calcium salt tetrahydrate (Sigma number P0378), α -p-Glucose-1-phosphate disodium salt hydrate (Sigma number PG4455), p-Glucose 6-phosphate sodium salt hydrate (Sigma number V900924), and *myo*-Inositol hexakisphosphate dipotassium salt (Sigma number P5681) in order. Spiked samples were extracted and analyzed by ³¹P NMR as described above.

2.5. Statistical analysis

Correlation analysis and data calculation were performed with SPSS 19 for Windows10 software (SPSS Inc., USA). The NMR spectra were processed with 10 Hz line broadening for the full spectra and 2 Hz line broadening to analyze the fine details of the orthophosphate monoester and diester regions. The spectra were processed using Bruker Software (TopSpin 3.5; Bruker Corporation; Germany). Corrections were made to account for phospholipid and RNA degradation by summing the peak areas associated with degradation products (α - and β -glycerophosphate and mononucleotides degraded from phospholipids and RNA, respectively), subtracting this sum from the α - and β -glycerophosphate and mononucleotide areas and adding them to the phospholipid and RNA areas (Cade-Menun, 2015).

3. Results

3.1. General sediment properties

Based on vertical profiles of TOC, TN, TOC/TN (weight ratio), δ^{13} C and δ^{15} N, the sediment can be approximately categorized into three distinct stages in evolution. In the section I (from -50 to 0 cm), TOC and TN gradually increased from 0.3% to 0.7% and from 0.04% to 0.07%, respectively, in an upward direction. The TOC/TN was < 8, but δ^{13} C and δ^{15} N were greater than -23% and 6.5%, respectively. In the section II (-50 cm to -150 cm), TOC gradually increased from 0.2% to 0.4%, but δ^{13} C and δ^{15} N decreased from -23% to -25% and 6.5% to 5‰, respectively. The weight ratio of TOC/TN increased from 4 to 10. In the section III (below -150 cm), the values of TOC, TN, TOC/TN, δ^{13} C and δ^{15} N kept narrow ranges. Overall, bulk isotope compositions of sediments displayed significantly positive shifts in δ^{13} C and δ^{15} N were both significantly negatively correlated with TOC/TN (δ^{13} C, $r^2 = 0.45$, Fig. S3b; δ^{15} N, $r^2 = 0.36$, Fig. S3c).

Concentrations of Al, Fe, and Mn presented the highest values in the upper 10 cm of the core, but the highest content of Ca was found near -20 cm. Distribution of Al below -20 cm was stable with the content of 34 g kg^{-1} . Fe gradually decreased from 32 g kg^{-1} to 22 g kg^{-1} between -50 cm and -100 cm, while increased from 22 g kg^{-1} to 30 g kg^{-1} between -100 cm and -150 cm. Fe was stable below -150 cm with the content of 30 g kg^{-1} . The content of Ca was stable at the depth between -50 cm and -150 cm, but gradually increased from 24 g kg^{-1} to 28 g kg^{-1} below -150 cm. Contents of Mn were low and stable below -20 cm. The EDS showed that content of siliceous compounds dominated the composition in the whole sediment core (Fig. S1).

3.2. Distributions of APA, Chl a and reactive Fe

Two indicators of biological reactivity (APA, Chl *a*) are shown in Fig. 4. APA fluctuated between $12 \text{ mg kg}^{-1} \text{ h}^{-1}$ and $57 \text{ mg kg}^{-1} \text{ h}^{-1}$, with an average of $24.7 \pm 13.5 \text{ mg kg}^{-1} \text{ h}^{-1}$. APA was high in the upper sediments (0 to -20 cm), then declined from -20 cm to -70 cm. APA increased again from -70 cm to -140 cm, and then remained stable below -150 cm. Chl *a* presented the highest value above -25 cm, with an average of $1.8 \pm 0.8 \text{ mg kg}^{-1}$. However, Chl *a* decreased from -18 to -35 cm, reached a minimum of $0.4 \pm 0.1 \text{ mg kg}^{-1}$ below -35 cm, then remained within a narrow range.

The vertical distributions of different active Fe are presented in Fig. 4. The average levels of Fe(II), Fe(III), and Fe $_{\rm RT}$ were 41.9 ± 15.8 µmol g⁻¹, 10.6 ± 7.3 µmol g⁻¹, and 52.5 ± 12.6 µmol g⁻¹, respectively. Both reactive Fe(II) and Fe $_{\rm RT}$ generally decreased above – 60 cm, increased markedly from – 60 to – 150 cm, then became stable, with means of 56.7 ± 2.8 µmol g⁻¹ and 58.3 ± 2.5 µmol g⁻¹, respectively, below – 150 cm. However, the reactive Fe(III) fluctuated in the upper section (from 0 to – 60 cm), decreased from – 60 to – 150 cm, and stabilized below – 150 cm. Additionally, reactive Fe(II) was significantly positive correlated with Fe $_{\rm RT}$ (r^2 = 0.78, Fig. S4a) and significantly negative correlated with Fe(III) (r^2 = 0.37, Fig. S4b).

3.3. Distributions of AVS, CRS, and ES in the sediments

As shown in Fig. 5, AVS varied from 12.2 to 40.3 mg kg^{-1} with depth. There was a high content of AVS in the upper 0 to -8 cm and at the bottom of the core (-310 to -370 cm). CRS and ES fluctuated from 21.3 to 1155.6 mg kg⁻¹ (average of 305.1 ± 328.7 mg kg⁻¹) and 8.2 to 87.1 mg kg⁻¹ (average of 30.8 ± 21.9 mg kg⁻¹), respectively. The CRS and ES concentrations remained high in the bottom section (-250 cm to

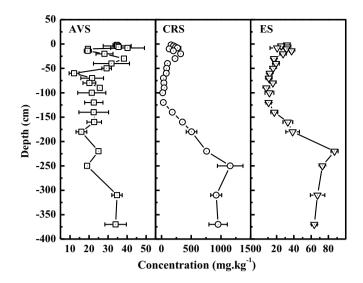


Fig. 5. Vertical distributions of AVS, CRS, and ES in the sediment profiles.

-370 cm) and low in the subsurface section (-20 cm to -100 cm) in the sediment core. In addition, RIS in the core was dominated by CRS (average of 65 \pm 19% of the total RIS), with AVS and ES accounting for an average of 24 \pm 16% and 11 \pm 4%, respectively.

3.4. Fractionation of P by the SMT and Ivanoff scheme

The results of P fractionation by the SMT protocol in the sediments shown in Fig. 6 (I) indicate that TP ranged from 478.6 mg kg⁻¹ to 552.6 mg kg⁻¹ (average of 508.1 ± 18.8 mg kg⁻¹). Ca–P was the dominant fraction of TP in the sediments, comprising 75% to 95% of the TP (average of 426.4 ± 20.6 mg kg⁻¹). OP was the second-most dominant TP fraction (10% to 15%). Al/Fe/Mn-P was the lowest TP component, accounting for 6% to 12%. The distributions of TP, Ca–P, Al/Fe/Mn-P, and OP by the SMT protocol in the sediment core can be divided into three general parts: they were relatively high and fluctuated in the surface sediment (0 to -20 cm), decreased with the depth ($-20 \sim -100$ cm) and remained mostly stable at lower depths. However, the decreasing trend of Ca–P was not obvious from -20 to -100 cm. The TP and Ca–P content increased slightly with depth below -100 cm.

The results in Fig. 6 (II) demonstrate that residue-OP and Flv-OP were the major fractions of OP. The residue-OP accounted for 45.8 \pm 9.6%,

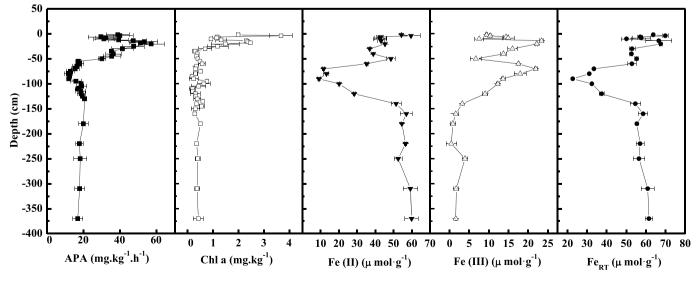


Fig. 4. Distributions of APA, Chl a, Fe(II), Fe(III) and Fe RT in the sediment profiles.

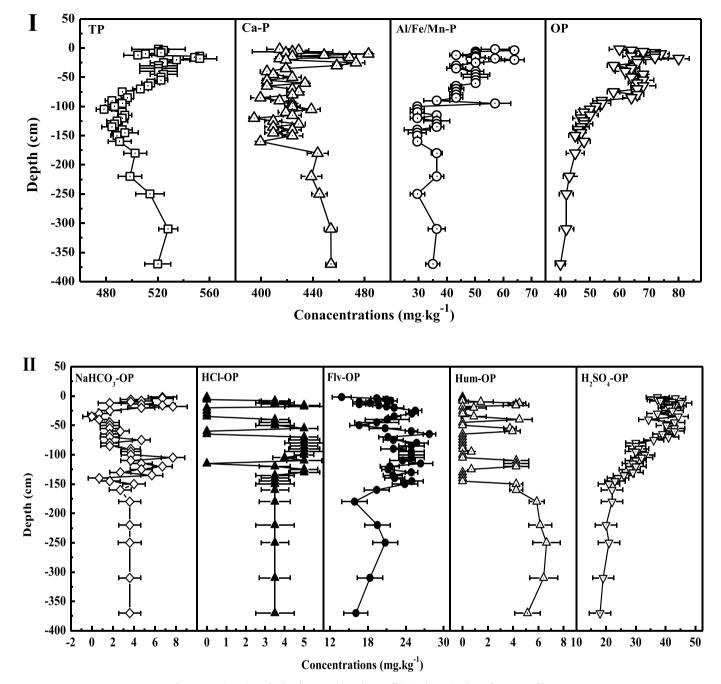


Fig. 6. Fractionation of P by the SMT (I) and Ivanoff (II) scheme in the sediment profiles.

and Flv-OP accounted for 35.8 \pm 6.6% of OP. The smallest fraction was Hum-OP accounting for 3.3 \pm 4.2% of the OP. The average residue-OP content was high in the upper section (above -50 cm). Subsequently, residue-OP dropped rapidly from 40 to 32 mg kg⁻¹ from -50 to -100 cm, but it decreased slowly from 32 to 18 mg kg⁻¹ from -100 to -370 cm. The Hum-OP concentrations were relatively low in the upper section (above -150 cm) but high in the deeper section (below -110 cm). Overall, Flv-OP was higher than Hum-OP.

3.5. ³¹P NMR spectra of NaOH-EDTA extracts of sediments

The NMR spectra of samples are shown in Fig. 7. Based on the spiking experiments, details of orthophosphate monoesters in sediments were identified and quantified (Fig. S2 and Table 1). The recoveries of TP and OP in NaOH-EDTA extracts were merely approximately 15% and 20%, respectively (Table S1). Therefore, there should be more P compounds that

could not be detected by ³¹P NMR. And the quantified content of various P in Table 1 was an underestimate result. Results in Table 1 indicate that IP mainly consisted of orthophosphate and OP was dominated by orthophosphate monoesters. Phospholipid and RNA were the dominant OP species in the sediments. The glucose-6-phosphate was detected in the surface sediment, and glucose-1-phosphate was in the bottom sediment.

4. Discussion

4.1. Eutrophication and source of OM in sediment

The TOC/TN weight ratio in sediments has been widely used to interpret OM source. TOC/TN has been reported to be 5–7 for marine algae and > 15 for terrestrial C3 vascular plants (Meyers, 1997). In this work, the TOC/TN ratios in the section I (above -50 cm) were closer to the range for typical autochthonous sources than for allochthonous

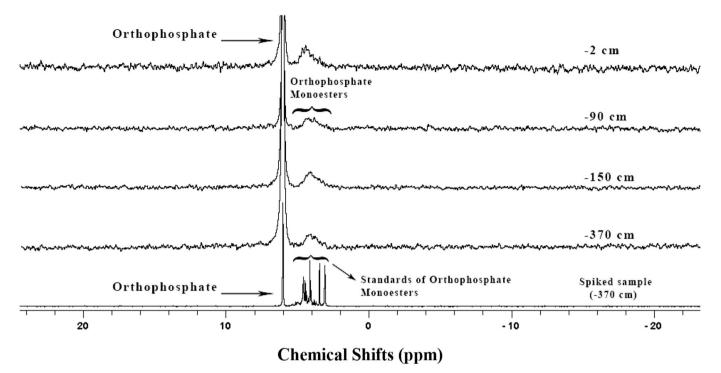


Fig. 7. ³¹P NMR spectra of NaOH-EDTA extracts from the sediment core.

sources, suggesting that OM in the section I of the sediment core was mainly derived from algal and bacterial production in the marine water column. However, the TOC/TN ratios were gradually trending towards the range for the terrestrial plants at the depth between -50 cm and -150 cm, which indicated that terrestrial OM inputs gradually became an important source below -50 cm. Unlike TOC/TN ratios, δ^{13} C and δ^{15} N values are not significantly influenced by sediment grain size and degradation of labile OM, making them useful in reconstructing past sources of OM in places with histories of changing depositional conditions (Meyers, 1997). Therefore, combined TOC/TN ratios and δ^{13} C and δ^{15} N values can more precisely distinguish between marine and continental plant sources of sedimentary OM. Phytoplankton in marine ecosystem usually have δ^{13} C values from -18 to -22% and the δ^{15} N value of dissolved NO_3^- used by marine algae is from 7 to 10‰. Terrestrial plants have δ^{13} C value from -25% to -33% and δ^{15} N value of atmosphere-derived N2 that is made available to land plants by nitrogen fixers in soil is approximately 0.5% (Fry and Sherr, 1984; Xiong et al., 2010). Thus, the above conclusion was further supported by δ^{13} C and δ^{15} N values. Based on the ²¹⁰Pb profiles recorded in sediments from the offshore of the North Yellow Sea (Wang et al., 2017), the sediment at -50 cm was accumulated during the 1970s. Therefore, variations in OM in sediment profiles indicated that there was a significant increase in the number of algae from the 1970s onwards, which suggested that the offshore sampling area of the Jiahe River estuary had entered a eutrophic or hyper-eutrophic stage in this period.

4.2. Redox-dependent vertical burial and regeneration of P

The PO_4^{3-} and phytoplankton debris in the overlying water reached to the seafloor easily and accumulated in the surface sediments. Additionally, Al, Fe and Mn contents were the highest in the surface sediments, resulting in a part of dissolved PO_4^{3-} in the overlying water to associate with the reactive sites on the surface of Fe/Al oxides and Mn oxides and convert into reactive Al/Fe/Mn-P. However, the other part of dissolved PO_4^{3-} could also be strongly associated with unreactive Al or Fe bound silicate (Al/Fe-Si) minerals (the unreactive Fe/ Al compounds) by Al–OH and Fe–OH functional groups inside the Al/ Fe-Si minerals through ion exchange (Kasama et al., 2004; Ruttenberg and Sulak, 2011). The PO_4^{3-} in Al/Fe-Si minerals is called occluded Al/ Fe-Si bound P (Al/Fe-Si-P), which is too stable to be released under natural conditions (Liu et al., 2003). A previous study indicated that α-Gly and β -Gly were degradation products from phospholipids and most mononucleotides originated from RNA or organisms (Cade-Menun, 2017). Settled phytoplankton debris contributed large amounts of OP in the surface sediments (Zhang et al., 2018). Surface sediments were aerobic because shallow seawater was enriched with oxygen, which could inhibit the reduction of Fe (III) oxides (hydroxides) and OP mineralization (Kraal et al., 2017). Therefore, higher concentrations of Al/ Fe/Mn-P and OP were preserved in the surface sediments. The highest content of Ca-P was found at around -20 cm just with the highest concentration of Ca, attributing to the dissolved PO_4^{3-} . The PO_4^{3-} can

Table 1

The relative distribution of P s	pecies determined by ³¹ F	P NMR spectroscopy in t	he NaOH-EDTA extracts.

Sediments	Integral area (%)									
	Ortho-P	myo-IHP	G6P	ChoP	α-Gly	β-Gly	AMP	G1P		
- 2 cm	92.3%	n.d.	2.26%	1.04	0.59%	1.08%	1.00	n.d.		
- 90 cm	91.7%	n.d.	n.d.	n.d.	0.63%	1.37%	1.17%	n.d.		
-150 cm	92.8%	n.d.	n.d.	0.72%	0.62%	1.24%	2.06%	1.23%		
- 370 cm	93.8%	n.d.	n.d.	n.d.	0.48%	0.87	0.86%	0.34%		

Note: n.d: not detected or below the LOQs. Ortho-P, orthophosphate; *myo*-IHP, *myo*-inositol hexakisphosphate; G6P, glucose-6-phosphate; ChoP, choline phosphate; G1P, glucose-1-phosphate.

be captured by the reactive Ca compounds or unreactive calcium-bound silicate transforming into stable biological apatite, CaCO₃ incorporated PO4³⁻, authigenic carbonate fluorapatite and/or occluded calciumbound silicate (Si-Ca-P) (Kang et al., 2017; Li et al., 2017). The PO₄³⁻ could be preserved within the Ca bound silicate minerals by OH functional groups inside the calcium-bound silicate through ion exchange (Kasama et al., 2004). Generally, the reactive Al/Fe/Mn-P would be easily released under sulfidic conditions (Li et al., 2016). In this study, the Al/Fe/Mn-P showed a sharp decrease from the upper sediment to the deep sediment. Most released Al/Fe/Mn-P should be reactive Al/Fe/ Mn-P. rather than occluded Al/Fe-Si-P. More specifically, redox-dependent P regeneration of reactive Al/Fe/Mn-P was the reactive Fe bound P (Fe-P). Previous studies indicated that Al bound P (Al-P) and Ca bound P (Ca-P) were essential inorganic components contributing to P retention in sediment, but they are not impacted by changes in redox potential (Alvarez-Rogel et al., 2007; Li et al., 2017). And the occluded Al/Fe-Si-P within the silicate lattice could avoid the regeneration under natural conditions (Matsue and Wada, 1989; Liu et al., 2003; Nasr et al., 2018; Zhu et al., 2018). Additionally, the residual Al/Fe/Mn-P below -100 cm was roughly constant, which could be stable as Al/Fe-Si-P and Al-P.

However, the Ca-P was relatively stable and difficult to be released for phytoplankton uptake (Qin and Zhu, 2006). Thus, Ca-P was a key burial phase in the sediments. Part of the PO₄³⁻ regenerated from the Fe-P and OP under reduced and alkaline phosphatase conditions could diffuse downward and adsorb onto clay or CaCO₃ to convert to authigenic carbonate apatite during early diagenesis (Ruttenberg and Berner, 1993). The other part of the released PO_4^{3-} could diffuse upward to be absorbed by the Al/Fe/Mn in the oxidizing condition (Zhu et al., 2018). Thus, there were two P "pumps" in the sediments. One was the Fe/Al/ Mn oxides in the upper sediment core; the other one was the Ca minerals in the lower sediment core, which could drive the released PO₄³⁻ to be buried as Al/Fe/Mn-P and Ca–P fractions. This resulted in the sharp increase in Al/Fe/Mn-P and Ca-P content from the middle of sediment core to the upper sediments and the lower sediments. Ca content gradually increased below -160 cm, leading to more PO_4^{3-} associated with reactive Ca compounds and/or Ca bound silicate, which could be another significant factor leading to the increase in Ca-P (apatite and/or Si-Ca-P) below -160 cm.

The recoveries of TP and OP in EDTA-NaOH extracts were both low mainly because high salinity reduced the extraction efficiency of EDTA-NaOH (Zhao et al., 2019b). And a high content of Ca-P and low content of TOC in sediment could also result in a lower recovery of P due to abundant unreactive P chelates firmly adhering to sediments (Zhu et al., 2018). Therefore, there should be more unreactive P compounds in the sediment core that could not be detected by ³¹P NMR because the recoveries of TP and OP were only approximately 15% and 20%, respectively. Ivanoff results further confirmed that there were more stable OP fractions and less reactive OP fractions in the marine sediment profile. A previous study indicated that some reactive OP (e.g., DNA, Glucose-1-phosphate, Glucose-6-phosphate, and phospholipid) chelated with metal ions or minerals to transform stable OP (Zhu et al., 2018). Generally, marine sediments contain abundant metal ions and unreactive minerals. Most of the reactive OP could chelate the humic acid and/or minerals to generate the recalcitrant OP chelates (Zhu et al., 2015). The OP metal chelates could be resistant to hydrolysis by phosphatase that could be stably preserved in sediments (Zhu et al., 2018). From the above, unreactive Al/Fe-Si-P, Al-P, Ca-P and OP metal chelates could be stably buried in anaerobic deep sediments, while free OP and reactive Al/Fe/Mn-P could be easily regenerated.

4.3. Sulfate reduction and pyritization promoted OP regeneration

The anaerobic decomposition of OP could act as electron donor during the Fe (III) reduction and sulfate reduction (Henneke et al., 1997), which promoted the regeneration of OP under anaerobic conditions.

There are two competitive pathways for Fe (III) reduction: one is microbial iron reduction (MIR) combined with OM oxidation (Eq. (1)), the other is chemical iron reduction (CIR) mainly by sulfide produced by sulfate reduction (Eq. (2)) (Canfield et al., 2005). Only when the sulfide is limited does MIR dominate in sediments (Wang and VanCappellen, 1996; Wijsman et al., 2002). In this study, the concentration of AVS, CRS and ES was low above -100 cm suggesting that the production of sulfide was limited. Therefore, MIR dominated Fe (III) reduction rather than CIR above -100 cm. AVS, CRS and ES gradually increased below -100 cm indicated that sulfides began to accumulate and the sulfate reduction gradually stronger. CIR would dominate the reaction instead of MIR under sufficient HS⁻ conditions below -100 cm. The enhanced reductive condition promoted sulfate reduction with HS⁻ accumulation in the deep sediments (Eq. (3)). Therefore, OP was decomposed as electron donor under the Fe (III) reduction that was mainly occurred above -100 cm, and OP decomposed as electron donor under the sulfate reduction was mainly occurred below -100 cm.

$$CH_2O + 4FeOOH + 8H^+ = CO_2 + 4Fe^{2+} + 7H_2O$$
 (1)

$$2FeOOH + 3H_2S + 4H^+ = 2FeS + S^0 + 4H_2O$$
(2)

$$SO_4^{2-} + 2CH_2O = H_2S + 2HCO_3^{-}$$
(3)

$$FeS + H_2S = FeS_2 \tag{4}$$

Previous studies reported that OP associated with metals was a stable complex form that was resistant to biotic or abiotic hydrolysis (Turner et al., 2012; Zhu et al., 2015). In this study, the accumulated HS⁻ from sulfate reduction below -100 cm could preferentially combine with metal ions including Fe(II) forming metal sulfides precipitation (Kraal et al., 2017), which reduced the efficiency of M^{n+} combining with OP generating stable OP- M^{n+} and/or OP- M^{n+} -OM complex (M^{n+} , metal ions). The free OP would be easily hydrolyzed by alkaline phosphatase after the metal ions were removed by HS⁻ (Zhu et al., 2015, 2018; Zhao et al., 2019a). Therefore, residue-OP dropped obviously with the accumulation of AVS and CRS below -100 cm. The formation of metal sulfides precipitation could promote the mineralization of OP. In addition, the free soluble HS^- is recognized as a biological toxicant that can poison cells, genes and enzymes due to its acidity and oxidative stress (Chen et al., 2019). And the toxicity of free soluble HS⁻ and metal ions to phosphatase has been reported by Zhao et al. (2019a). Therefore, many trace metal ions and freely soluble HS⁻ can be collectively eliminated by the formation of metal sulfides precipitations (Jong and Parry, 2003). And the inhibited phosphatase activity could be restored after the elimination of free soluble HS⁻ and metal ions (Zhao et al., 2019a). Therefore, APA increased with the formation of AVS and CRS below -100 cm. The formation of insoluble metal sulfides precipitation promoted the release of PO₄³⁻ from OP. Therefore, sulfate reduction and pyritization promoted partial stable OP regeneration.

5. Conclusion

The indigenous biological input was the primary source of OM in the upper sediments. Most released Al/Fe/Mn-P under reducing conditions should be reactive Fe–P, rather than unreactive Al/Fe-Si-P and Al–P. The regeneration of $PO_4{}^{3-}$ from free OP is the sediment core mainly caused by the sulfate reduction and hydrolysis by phosphatase. The released $PO_4{}^{3-}$ from Fe–P and OP could be captured by the Al/Fe/ Mn (oxyhydr) oxides and Ca minerals to transform into Al/Fe/Mn-P and Ca–P. The Ca–P, Al–P, unreactive Al/Fe-Si-P and some of stable metal chelated OP were the main burial fraction of P. However, sulfate reduction and formation of insoluble metallic sulfides including pyritization could promote partial non-labile OP decomposition by anaerobic decomposition, removing metal ions from the "Mⁿ⁺-OP" chelates and restoring phosphatase activity.

Acknowledgments

This study was supported by the Strategic Priority Research Program of the Chinese Academy of Sciences (Grant No.:XDA23050203) and National Natural Science Foundation of China (Grant No.:41373100). Additional support was provided by the Key Project of Research and Development Plan of Yantai (Grant No.: 2018ZHGY083) and the Key Research and Development Program of Shandong Province (Grant No. 2019GSF109002).

Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.marpolbul.2019.110582.

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